# Final Report: 2006 and 2007 Data

# Concentration and Effects of Selenium in California Gulls Breeding on the Great Salt Lake

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# Abstract

We examined selenium concentrations in California gulls (Larus californicus) nesting on the Great Salt Lake, Utah during 2006 and 2007. During 2006, the mean selenium concentration  $(\pm SE)$  in adult blood samples was  $18.1 \pm 1.5 \,\mu g/g$  (n = 35) on a dry weight basis,  $8.1 \pm 0.4$  in adult liver samples (n = 36), and  $3.0 + 0.10 \mu g/g$  in eggs (n = 35). During 2007, selenium concentrations were  $15.7 \pm 1.5 \,\mu\text{g/g}$  in blood and  $8.3 \pm 0.4 \,\mu\text{g/g}$  in liver; mercury concentrations were  $2.4 \pm 0.3$  $\mu g/g$  in blood and  $4.1 \pm 0.5$  in liver. Body mass was not correlated with selenium or mercury concentrations in the blood or liver for either adult males or females. Gulls collected from different Great Salt Lake colonies varied significantly in the concentration of selenium in their blood but not in livers or eggs. Selenium concentrations were higher in blood of gulls collected at the GSLM colony than in gulls collected from the Antelope Island colony or Hat Island colony. Gulls collected from a freshwater colony (Neponsett Reservoir) located in the headwaters of the Bear River had similar levels of selenium in the blood and liver as gulls collected on the Great Salt Lake but lower mercury levels. Of 72 eggs collected at random from Great Salt Lake colonies, only one showed no embryo development, and none of the embryos exhibited signs of malposition or deformities. We examined 100 newly hatched chicks from Great Salt Lake colonies for teratogenesis; all chicks appeared normal. Hence, the high selenium concentrations in blood of adult gulls do not seem to be impairing the gulls' health or reproductive ability.

# Introduction

Selenium is a naturally occurring trace element, and small quantities of it are essential for animal health. However, it becomes toxic at higher concentrations. High concentrations of selenium have been shown, both in captive and free-ranging birds, to cause reduced egg hatchability, embryonic defects, and lower survival rates of chicks and adults (Ohlendorf et al.1989, Ohlendorf 2003). For example, birds foraging in California's Kesterson Reservoir, which was the disposal site for subsurface agricultural drainage from portions of the western San Joaquin Valley, accumulated high concentrations of selenium in their tissues (Ohlendorf 2002, 2003). Selenium concentrations in eggs of all aquatic birds collected from Kesterson Reservoir were higher than background levels (3  $\mu$ g/g), and eared grebes (*Podiceps nigricollis*) had the highest concentrations, with a mean of 70  $\mu$ g/g (all weights are reported on a dry-weight basis unless otherwise noted). Elevated concentrations of selenium impaired the reproductive ability of several avian species nesting at Kesterson Reservoir and caused mortality of adult birds (Ohlendorf et al. 1989; Ohlendorf 2002, 2003).

Birds accumulate selenium from their food, and if they consume a diet too rich in selenium during the pre-laying period, they transfer selenium to their eggs, causing harmful effects. In laboratory studies, dietary concentrations of about 4.9  $\mu$ g Se/g of food reduced the reproductive success of mallards (*Anas platyrhynchos*; Ohlendorf 2003). Mallard ducklings maintained on a diet containing 40  $\mu$ g/g or greater selenium concentrations exhibited high mortality rates within 6 weeks of starting the diet. At Kesterson Reservoir, where many avian species that exhibited reproductive problems, some aquatic insects had mean selenium concentrations greater than 100  $\mu$ g/g (Hothem and Ohlendorf 1989, Schuler et al. 1990).

Higher selenium concentrations also can impair the health of adult birds. Mallards maintained on a diet enriched with  $20~\mu g$  Se/g of food had lesions in their liver and integument. Mallards on a diet of  $40~\mu g/g$  lost weight and exhibited abnormalities such as feather loss, loss of nails, and beak necrosis (Albers et al. 1996, O'Toole and Raisbeck 1998).

The Great Salt Lake (GSL) in Utah is an important habitat for many avian species. For instance, about half of the world's eared grebes spend the fall there eating brine shrimp and accumulating enough nutrients to fly to their wintering grounds in Mexico and California. During the breeding season, the lake also provides foraging and nesting habitat for California gulls (*Larus californicus*) and shorebirds. Hence, there is a need to ensure that selenium concentrations in the GSL do not reach concentrations that would impair the health or reproduction of the birds that depend upon the GSL. For this reason, the Utah Division of Water Quality wants to establish a water standard for selenium in the GSL. To aid this effort, we measured selenium concentrations in adult California gulls breeding in three different parts of the GSL and in their eggs. We sampled their food during the pre-laying period and took water and sediment samples in their feeding grounds to assess them for selenium. We also checked California gull eggs for viability and embryos for deformities. This study was designed to answer the following specific questions.

- 1. What is the diet of California gulls during the egg-laying period?
- 2. What are the ambient selenium concentrations in the water, sediment, and diet items at the foraging sites of nesting California gulls in the GSL?
- 3. Are selenium and mercury levels in gulls nesting in the saline environment of the GSL similar to gulls nesting on a freshwater reservoir?
- 4. What is the relationship between selenium and mercury concentrations in the liver and blood of adult gulls and eggs?
- 5. What are the associated selenium concentrations in nesting California gulls (blood and liver), a random sample of gull eggs, and gull eggs that failed to hatch?

#### Methods

Collection of adult gulls. During 2006, we collected gulls from nesting colonies located on the Great Salt Lake at Hat Island, the Great Salt Lake Mineral (GSLM) facility, Egg Island, and White Rock Island. Egg Island and White Rock Island are small rocky islands within a few kilometers of each other. Both are within 1 km of the much larger Antelope Island. The gulls nesting in these 2 colonies use the same foraging areas (Conover, personal observation). Hence, we considered Egg Island and White Rock Island as a single colony and referred to them as the Antelope Island colony (Figure 1). During 2007, we re-sampled gulls from Hat Island and GSLM colonies and also collected California gulls from a Neponsett Reservoir colony, which is located in the headwaters of the Bear River in Rich County, Utah.

During the period when the gulls were laying eggs, we used a shotgun to collect 12 gulls from three GSL colonies (Hat Island, GSLM, and Antelope Island). To do this, we positioned ourselves 0.5-1.0 km from a colony and shot gulls that were flying back to the colony in a straight line. We assumed that these gulls were more likely to have food in their esophagus than gulls leaving the colony or gulls that were flying slowly in a circular pattern and appeared to be searching for food. All gulls from a single colony were shot on the same day. We immediately used a syringe to collect at least 1 ml of blood from the thoracic cavity. The blood was kept in the syringe and frozen. Within 12 hours of when the birds were shot, we collected all food from the bird's esophagus and obtained a liver sample. The liver sample was placed in a Whirl-Pak® bag and frozen. Esophagus samples were weighed (wet weight) and were stored in 95% alcohol. We weighed and measured the birds, determined their age by examining their plumage, and their sex by examining the gonads. Food in the esophagus was weighed and sorted by content or species. We determined the proportion of a food sample that could be attributed to different types of food or species.

Before we started collecting gulls, we observed where the gulls were foraging. We then went to those foraging sites and collected food samples by dragging a 1-m<sup>2</sup> circular seining net with a 5-mm mesh size behind the boat at a speed that just kept the top of the net at the top of the water. Hence, all food samples were obtained from the upper 1 m of the water column. Five seine samples were collected at each colony. Each of these was placed in a separate Whirl-Pak® bag and frozen.

We also collected 5 water samples from the upper 1 m of the water column and used a core to obtain 5 sediment samples from the top 0.1 m of the bottom sediment. The water and sediment samples were collected at the same places where the food samples were collected. These were placed in polypropylene vials and maintained at room temperature. Equal volumes of water from each of five water samples collected at the foraging grounds near each colony were combined to create a single composite water sample per colony. Likewise, the five sediment samples were combined in equal volume and homogenized to create a composite sediment sample for each colony

Selenium Analysis. Blood and liver samples from 11 or 12 gulls at each colony were sent to Laboratory and Environmental Testing Incorporated (LET), Columbia, Missouri, for selenium analysis. LET analyzed the tissue for total selenium using hydride generation atomic absorption spectrometry, with a target reporting limit of  $0.2~\mu g/g$ . Quality control of selenium analyses was conducted using one or more method blanks, matrix spikes, matrix spike duplicates, and reference samples for each sample batch (CH2M HILL 2006). The seine nets collected almost pure samples of brine shrimp, and the five brine shrimp samples from each colony were also sent to LET for selenium analysis. The water samples were sent to Frontier Geoscience, Seattle, Washington, and sediment samples were sent to LET for selenium analysis.

Collection of California gull eggs. We collected a single egg from each of 24 nests in each GSL colony (72 eggs total) when approximately 10% of the nests contained a chick or pipped egg. (This assured that the eggs we collected were likely to have late-stage embryos; therefore all [or almost all] eggs contained embryos assessable for developmental abnormalities.) All eggs were collected from three-egg clutches where no eggs had hatched. These nests were selected by placing a conceptual gird over the colony containing a series of numbered points, selecting points from a random numbers tables and sampling the nest located closest to that point that met the criteria. Which of the three eggs in that nest was collected was determined by numbering each egg in the clutch and selecting which number to sample based on a random numbers tables. All eggs were collected during 2006 except for the Neponsett Reservoir colony eggs, which were collected during 2007.

Eggs were stored in a refrigerator, and embryos were examined within four days of collection when samples were being prepared for selenium analysis. Each embryo was checked for stage of

embryonic development (embryo age) by comparing to existing aging criteria and atlases (Hamburger and Hamilton 1951; Hamilton 1952; Pisenti et al. 2001) and developmental abnormalities, including a determination of the embryo's pre-hatch position in the egg (i.e., for malposition) based on Romanoff and Romanoff (1972). An egg was considered viable if it contained a developing late incubation stage embryo. The contents of each egg (including the embryo) were placed in a marked chemically-cleaned container and preserved frozen for later chemical analysis. Eleven or 12 eggs from each colony were analyzed for total selenium by LET, and the others were stored for possible later analysis.

Examination of newly hatched chicks of California gulls and salvaged eggs for deformities. Immediately after the chicks hatched, we revisited the GSLM, Hat, and Antelope colonies to check 100 chicks that had hatched within the last 12 hours for deformities. Forty-eight salvaged eggs also were collected from the Hat Island and GSLM colonies (24 from each colony). A salvaged egg was defined as an egg remaining in a nest after the other eggs in the nest had hatched and that was no longer being incubated (i.e., egg was at ambient temperature). Salvaged eggs were checked to determine fertility and the presence of dead embryos. All embryos (including all contents of those eggs) were placed in chemically-cleaned containers and preserved frozen for later analysis.

Statistical analyses. Data on selenium concentrations were normally distributed based on the D'Agostino-Pearson Omnibus  $K^2$  normality test. Hence, parametric statistics were used. Correlations were conducted to compare selenium concentration in an individual gull's blood and liver. The same analyses were used to compare these to the gulls' mass and mercury concentrations in their blood and liver. Unpaired Students t-tests or F-tests were used to test for differences in mass and selenium concentrations. F-tests were used to test if selenium concentrations differed among colonies. In all tests, results were considered significant if P < 0.05.

# Results

Food analyses for adults.—Thirty of the 35 adult gulls collected during 2006 had food in their esophagus (Appendix 1). Only one gull had more than a single kind of food item in its esophagus. That one contained 60% brine shrimp, 35% corixids, and 5% adult midges. For the 29 gulls that contained a single food item, 21 (75%) contained brine shrimp, 2 (7%) corixids, 2 (7%) brine fly larvae, 1 (4%) hot dogs, 1 (4%) earthworms, and 1 (4%) rotten carp (*Cyprinus carpio*) flesh. At all colonies, most gulls contained only brine shrimp. Corixids and midges were detected only at the GSLM colony. The earthworm and carp samples came from Antelope Island colony; hot hogs came from Hat Island colony.

Thirty two of the 36 adult gulls collected during 2007 had food in their esophagus (Appendix 2). Three gulls had more than a single kind of food item in its esophagus and those three had a combination of food from terrestrial sources (i.e., garbage and insects). Six gulls from GLSM colony contained brine shrimp, 4 had midge larvae, and 2 contained garbage. Ten of 12 gulls from Hat Island had eaten brine shrimp exclusively, and the other two contained either garbage or terrestrial insects in their esophagus. The eight gulls from Neponsett Reservoir that had food in their esophagus had fed on garbage and terrestrial insects.

Selenium analyses of adults collected during 2006.—Among individual gulls, selenium concentrations in blood and liver were highly correlated ( $r^2 = 0.78$ , F = 117.22; d.f. = 1, 32; P = 0.0001 [Figure 2]). There was no significant difference (t = 1.56, d.f. = 27, P = 0.13) between the selenium concentrations (mean  $\pm$  SE) in the livers of adult males ( $7.4 \pm 0.5 \,\mu\text{g/g}$ ) and adult females ( $8.7 \pm 0.8 \,\mu\text{g/g}$ ). Likewise, there was no significant difference (t = 1.75, d.f. = 27, P = 0.09) between selenium concentrations (mean  $\pm$  SE) in the blood of adult males ( $15.2 \pm 1.6 \,\mu\text{g/g}$ ) and adult females

 $(20.6 \pm 3.0 \,\mu\text{g/g})$ . Hence, data from the two sexes were combined for further analyses. For all adults combined, selenium concentrations in blood samples were  $18.1 \pm 1.5 \,\mu\text{g/g}$  (n = 35); they were  $8.1 \pm 0.4$  in liver samples (n = 36; Appendix 2).

There was no significant difference in blood selenium concentrations (F = 0.34; d.f. = 1, 27; P = 0.56) between the 22 gulls that had mainly brine shrimp in their esophagus ( $16.9 \pm 1.8 \,\mu\text{g/g}$ ) and the 7 gulls that had other types of food in their esophagus ( $19.2 \pm 4.0 \,\mu\text{g/g}$ ). Likewise there was no significant difference (F = 0.12; d.f. = 1, 27; P = 0.73) between selenium levels in the liver of gulls that fed on brine shrimp ( $8.4 \pm 1.1 \,\mu\text{g/g}$ ) and those that fed on some other type of food ( $8.0 \pm 0.5 \,\mu\text{g/g}$ ).

Among gulls collected from different colonies, there was a significant difference in the concentration of selenium in blood (F = 6.27; d.f. = 2, 32; P = 0.005) but not in livers (F = 1.85; d.f. = 2, 32; P = 0.17) (Table 1). Selenium concentrations were highest in blood of gulls collected at the GSLM colony, which is close to where water from the Bear River flows into GSL, and lowest in gulls from Antelope Island colony. Gulls from Hat Island colony had intermediate concentrations of selenium. This pattern of the highest selenium concentrations being recorded at the GSLM colony was true for selenium concentrations in blood, liver, eggs, and sediment although differences among colonies were significant only for blood

Not surprisingly, there was a significant difference (F = 10.31; d.f. = 1, 26; P = 0.004) in the body mass of males ( $727 \pm 16.4$  g) and females ( $628 \pm 23.2$  g). Hence, the effects of selenium on body mass were analyzed separately for each sex. For males, body mass was not correlated with selenium concentrations in blood ( $r^2 = 0.01$ , F = 0.15; d.f. = 1, 15; P = 0.71) or liver ( $r^2 = 0.002$ , F = 0.00; d.f. = 1, 15; P = 0.96 [Figure 3]). Likewise for females, body mass was not correlated with selenium concentrations in blood ( $r^2 = 0.01$ , F = 0.78; d.f. = 1, 9; P = 0.40) or liver ( $r^2 = 0.03$ , F = 0.23; d.f. = 1, 15; P = 0.64 [Figure 4]).

Selenium and mercury analyses of adults during 2007. For all adults collected during 2007 (n = 36), selenium concentrations were  $15.7 \pm 1.5 \,\mu\text{g/g}$  in blood and  $8.3 \pm 0.4$  in liver (Appendix 2). For these same birds, mercury concentrations were  $2.4 + 0.3 \,\mu\text{g/g}$  in blood and 4.1 + 0.5 in liver.

Among individual gulls, selenium concentrations in blood and liver were highly correlated ( $r^2 = 0.70$ , F = 80.79; d.f. = 1, 34; P = 0.001) as was mercury concentrations in blood and liver ( $r^2 = 0.74$ , F = 95.03; d.f. = 1, 34; P = 0.001). Blood selenium concentrations were correlated with mercury levels in blood ( $r^2 = 0.14$ , F = 5.75; d.f. = 1, 34; P = 0.02) but not mercury levels in liver ( $r^2 = 0.05$ , F = 1.85; d.f. = 1, 34; P = 0.18). Selenium concentrations in liver were not correlated with either mercury levels in the blood ( $r^2 = 0.07$ , F = 2.52; d.f. = 1, 34; P = 0.12) or liver ( $r^2 = 0.03$ , F = 1.22; d.f. = 1, 34; P = 0.28).

Among gulls collected during 2007, the highest selenium concentrations were once again found in adult gulls and eggs collected from GSLM colony (Table 2). In fact, selenium levels in GSLM gulls were significantly higher than those gulls from Hat Island but not from Neponsett gulls, which had intermediate levels of selenium (Table 2). Neponsett gulls had intermediate levels of selenium. When gulls collected at GSLM colony during 2007 were compared to those collected during 2006 (Tables 1 and 2), blood selenium concentrations were similar (F = 0.78; d.f. = 1, 21; P = 0.39) as were liver selenium levels (F = 0.00; d.f. = 1, 21; P = 0.95). For gulls collected at GSLM colony, those collected during 2006 had higher selenium levels in their blood than those collected during 2007 (F = 4.57; d.f. = 1, 22; P = 0.04) but selenium levels in their livers were similar (F = 0.59; d.f. = 1, 22; P = 0.59).

Mercury concentrations in blood and liver were similar in gulls collected from Hat Island and GSLM colonies (Table 2). However gulls from the freshwater colony (Neponsett Reservoir) had significantly lower mercury concentrations in blood and liver than gulls from Hat Island and GLSM colonies (Table 2).

Effects of selenium and mercury on body mass were analyzed separately for each sex. For males, body mass was not correlated with selenium concentrations in blood ( $r^2 = 0.01$ , F = 0.15; d.f. = 1, 15; P = 0.71), selenium concentrations in liver ( $r^2 = 0.002$ , F = 0.00; d.f. = 1, 15; P = 0.96), mercury concentrations in blood ( $r^2 = 0.01$ , F = 0.15; d.f. = 1, 15; P = 0.71), or mercury concentrations in liver ( $r^2 = 0.002$ , F = 0.00; d.f. = 1, 15; P = 0.96). Likewise for females, body mass was not correlated with selenium concentrations in blood ( $r^2 = 0.01$ , F = 0.78; d.f. = 1, 9; P = 0.40), selenium concentrations in liver ( $r^2 = 0.03$ , F = 0.23; d.f. = 1, 15; P = 0.64), mercury concentrations in blood ( $r^2 = 0.01$ , F = 0.15; d.f. = 1, 15; P = 0.71), or mercury concentrations in liver ( $r^2 = 0.002$ , F = 0.00; d.f. = 1, 15; P = 0.96).

*Selenium and mercury analyses of food.*–During 2006, selenium concentrations in water and brine shrimp were highest at the Hat Island colony (Table 1). For the water and sediment samples, only a single sample was analyzed from each colony, and statistics could not be used to test these variables.

During 2007, selenium concentrations in brine shrimp at Hat Island were once again higher than at GSLM colony, but mercury levels were similar (Table 2). Mercury concentrations in brine shrimp from the two colonies were similar. Brine shrimp collected at Hat Island during 2006 contained higher selenium concentrations than samples collected from the same colony during 2007 (F = 27.09; d.f. = 1, 8; P = 0.001). Likewise, brine shrimp collected from GSLM colony during 2006 had higher selenium levels than those collected during 2007 (F = 13.83; d.f. = 1, 8; P = 0.006). Food samples from Neponsett Reservoir colony were not analyzed because most gulls were foraging on bread and garbage and there seemed little need to determine the selenium or mercury concentration of bread.

Selenium and mercury analyses of eggs.—Selenium concentrations in eggs collected randomly during 2006 were  $3.0 \pm 0.10 \,\mu\text{g/g}$  (n = 35). Selenium concentrations did not differ (F = 1.76; d.f. = 2, 32; P = 0.19) among eggs collected from the different GSL colonies (Table 1 and Appendix 3).

Eggs collected randomly from Neponsett Reservoir during 2007 had selenium concentrations of  $2.8 \pm 0.10 \,\mu\text{g/g}$  and mercury concentrations of  $0.26 \pm 0.05 \,\mu\text{g/g}$  (n = 12). Selenium concentrations for eggs collected at Neponsett Reservoir differed from those collected at the GSLM colony (F = 8.31; d.f. = 1, 21; P = 0.009) but not from eggs collected at Hat Island (F = 0.03; d.f. = 1, 21; P = 0.87) or Antelope Island (F = 0.01; d.f. = 1, 21; P = 0.92). For eggs collected at Neponsett Reservoir, selenium concentrations were not correlated with mercury concentrations ( $r^2 = 0.03$ ; F = 0.30; F = 0.30; F = 0.30; F = 0.60).

Analyses of eggs and chicks for viability and deformities.—Among the sample of 24 eggs randomly sampled from 3-egg clutches during the late incubation period from GSL colonies (72 eggs total), all contained developing late incubation stage embryos except a single egg that came from the GSLM colony (Appendix 3). None of the embryos exhibited signs of malposition or deformities. We examined 100 newly hatched chicks from GSL colonies for teratogenesis; all chicks appeared normal. Out of 48 salvaged eggs from GSL colonies, 38 contained dead embryos; all embryos were normal in appearance and position.

During 2007, 1 of 12 eggs collected at Neponsett Reservoir colony was rotten, and one had no embryo (Appendix 3). Ten eggs contained late incubation stage embryos, and none of the embryos exhibited signs of malposition or deformities.

#### Discussion

In California gulls, we found that selenium concentrations ranged from 4 to 15  $\mu$ g/g in livers. Mean background selenium concentrations have been reported to be <10  $\mu$ g/g in livers (USDI 1998, Ohlendorf 2003). We detected selenium concentrations in California gull eggs ranging from 2.0 to 4.3  $\mu$ g/g in eggs. Mean background selenium concentrations for individual eggs are considered to be < 5  $\mu$ g/g (USDI 1998, Ohlendorf 2003) or < 3  $\mu$ g/g for population means (Skorupa and Ohlendorf 1991). Hence, selenium concentrations in our egg and liver samples were generally consistent with background concentrations.

Surprisingly, selenium concentrations in blood of gulls nesting on GSL ranged from 5 to  $46~\mu g/g$ . These concentrations were higher than we expected given the concentrations found in livers, eggs, and diets. In selenium feeding studies of mallards (*Anas platyrhynchos*; Heinz and Fitzgerald 1993) and American kestrels (*Falco sparverius*; Yamamoto et al. 1998), blood selenium concentrations did not significantly exceed dietary concentrations and were similar to diet concentrations after four to eight weeks. We found that mean selenium concentrations in the blood of gulls from different GSL colonies were 2.4 to 5.5 times higher than selenium concentrations in the brine shrimp upon which they were foraging.

Selenium concentrations in the blood of predatory terrestrial birds (kestrel, red-tailed hawk [Buteo jamaicensis], northern harrier [Circus cyaneus], barn owl [Tyto alba], and loggerhead shrike [Lanius ludovicianus]) from a contaminated grassland in California ranged from 1.5 to 38  $\mu$ g/g dry weight (Santolo and Yamamoto 1999). Selenium concentrations in whole blood above 2  $\mu$ g/g dry weight are considered to exceed normal background, and 5  $\mu$ g/g dry weight is considered a provisional threshold indicating that further study is warranted (USDI 1998). However, toxicity studies of gulls were not reviewed for the development of those guidelines, and the ecotoxicology of selenium to gulls may differ from that for other species. Interestingly, we found that California gulls collected at a freshwater colony (Neponsett Reservoir) had selenium levels in their blood similar to those of GSL gulls but lower mercury concentrations. These results suggest that high selenium concentrations in blood may be a species trait rather than a characteristic of a saline environment.

Reasons for the anomalously high selenium concentrations in blood, but much lower concentrations in liver and eggs are not known. A possible explanation for the elevated concentrations of selenium in our blood samples may be relatively high mercury concentrations found in the Great Salt Lake ecosystem. Selenium and mercury may interact to form a stable, nontoxic complex so that selenium may provide adult birds some protection from mercury toxicity (Ohlendorf 2003, Wiener et al. 2003). This interaction between mercury and selenium may cause an enhanced accumulation and retention of both chemicals in birds (Furness and Rainbow 1990, Scheuhammer et al. 1998, Spalding et al. 2000, Henny et al. 2002). Differences in blood and liver concentrations of selenium may result from faster selenium elimination in liver than blood and to the binding of selenium to inorganic mercury creating an inert mercury-selenium protein (Wayland et al. 2001). In wading birds, selenium and mercury concentrations were positively correlated in the blood, but not in liver or kidney tissues (Goede and Wolterbeek 1994).

Although the few studies of selenium-mercury interaction in birds used various forms of Se and Hg (some not using environmentally relevant forms), they do provide approximations of potential effects. In a study by Heinz and Hoffman (1998) using mercury as methylmercury chloride

and seleno-DL-methionine, captive female mallards fed a diet containing both 10  $\mu$ g Se/g of feed and 10  $\mu$ g Hg/g had a selenium concentration in the liver 1.5 times higher than females fed a diet containing just selenium (10  $\mu$ g Se/g). In the same experiment, male mallards fed the selenium and mercury combination diet had almost 12 times the selenium concentration of male mallards fed the selenium-only diet. Similar results were found with Japanese quail fed diets containing methylmercury and selenite (El-Begearmi et al. 1977, 1982). However, our results suggest that a selenium-mercury interaction may not be responsible for the high selenium levels in California gulls. Among individual gulls, we found a statistically significant but weak correlation ( $r^2$  = 0.14) between the concentrations of selenium and mercury in blood but no correlation between selenium levels in blood and mercury levels in liver. Also, gulls from Neponsett Reservoir had similar selenium concentrations in their blood as GSL gulls, but they had much lower mercury concentrations.

Among free-ranging birds, sensitivity to selenium varies among species. In black-necked stilts (*Himantopus mexicanus*), the threshold for teratogenesis (EC<sub>10</sub>) was 37  $\mu$ g/g in eggs (Skorupa 1998, Ohlendorf 2003). However, the EC<sub>10</sub> was 23  $\mu$ g/g for mallards and 74  $\mu$ g/g for American avocets (*Recurvirostra americana*). Even lower concentrations of selenium can cause a decrease in egg viability. Selenium concentrations in eggs as low as 6–7  $\mu$ g/g resulted in reduced viability of eggs in black-necked stilts. Heinz (1996) suggested that 10  $\mu$ g/g be considered the threshold where selenium concentrations start to have an effect on the hatchability of bird eggs, while Fairbrother et al. (1999) recommended a threshold concentration of 16  $\mu$ g/g.

We found selenium concentrations in 30 California gull eggs collected from GSL colonies ranged from 2.0 to 4.3  $\mu$ g/g. These concentrations were similar to California gulls eggs collected from Neponsett Reservoir colony located in the upper watershed of the Bear River and are below the concentrations shown in other avian species to cause teratogenesis or a significant decrease in egg viability. We detected no evidence that these concentrations of selenium were causing an adverse effect on California gulls nesting on GSL.

# References

- Albers, P. H., D. E. Green, and C. J. Sanderson. 1996. Diagnostic criteria for selenium toxicosis in aquatic birds: dietary exposure, tissue concentrations, and macroscopic effects. *Journal of Wildlife Diseases* 32:468–485.
- CH2M HILL. 2006. Development of a selenium standard for the open waters of the Great Salt Lake. Utah Department of Environmental Quality, Division of Water Quality, Salt Lake City, Utah.
- El-Begearmi, M. M., M. L. Sunde, H. E. Ganther. 1977. A mutual protective effect of mercury and selenium in Japanese quail. *Poultry Science* 56:313–322.
- El-Begearmi, M. M., H. E. Ganther, M. L. Sunde. 1982. Dietary interaction between methylmercury, selenium, arsenic, and sulfur amino acids in Japanese quail. *Poultry Science* 61:272–279.
- Fairbrother, A., K. V. Brix, J. E. Toll, S. McKay, and W. J. Adams. 1999. Egg selenium concentrations as predictor of avian toxicity. *Human and Ecological Risk Assessment* 5:1229–1253.

Furness, R. W., and P. S. Rainbow. 1990. *Heavy Metals in the Marine Environment*. CRC Press, Boca Raton, Florida.

- Goede, A. A., and H. T. Wolterbeek. 1994. Have high selenium concentrations in wading birds their origin in mercury. *Science of the Total Environment* 144:247–253.
- Hamburger, V., and H. L. Hamilton. 1951. A series of normal stages in the development of the chick. *Journal of Morphology* 88:49-92.
- Hamilton, H. L. 1952. Lillie's Development of the Chick. Henry Holt and Co., Inc, New York, New York
- Heinz, G. H. 1996. Selenium in birds. Pages 447–458 in W. N. Beyer, G. H. Heinz, and A. W. Redmon-Norwood, editors. *Interpreting Environmental Contaminants in Animal Tissues*. Lewis Publishers, Boca Raton, Florida.
- Heinz, G. H. and D. J. Hoffman. 1998. Methylmercury chloride and selenomethionine interactions on health and reproduction in mallards. *Environmental Toxicology and Chemistry* 17:139-145.
- Henny, C. J., E. F. Hill, D. J. Hoffman, M. G. Spalding, and R. A. Grove. 2002. Nineteenth century mercury: hazard to wading birds and cormorants of the Carson River, Nevada. *Ecotoxicology* 11:213–231.
- Hothem, R. L., and H. M. Ohlendorf. 1989. Contaminants in foods of aquatic birds at Kesterson Reservoir, California, 1985. *Archives of Environmental Contamination and Toxicology* 18:773–780.
- Ohlendorf, H. M. 2002. The birds of Kesterson Reservoir: a historical perspective. *Aquatic Toxicology* 57:1–10.
- Ohlendorf, H. M. 2003. Ecotoxicology of selenium. Pages 465–500 *in* D. J. Hoffman, B. A. Rattner, G. A. Burton, Jr., and J. Cairns, Jr. *Handbook of Ecotoxicology*. Lewis Publishers, Boca Raton, Florida.
- Ohlendorf, H. M., R. L. Hothem, and D. Welsh. 1989. Nest success, cause-specific nest failure, and hatchability of aquatic birds at selenium-contaminated Kesterson Reservoir and a reference site. *Condor* 91:787–796.
- O'Toole, D., and M. R. Raisbeck. 1998. Magic numbers, elusive lesions: comparative aspects of selenium toxicosis in herbivores and waterfowl. Pages 335–395 in W. T. Frankenberger, Jr., and R. A. Engberg, editors. *Environmental Chemistry of Selenium*. Marcel Dekker, New York, New York.
- Pisenti, J. M., G. M. Santolo, J. T. Yamamoto, and A. A. Morzenti. 2001. Embryonic development of the American kestrel (*Falco sparverius*): external criteria for staging. *Journal of Raptor Research* 35:194-206.
- Romanoff, A. L., and A. J. Romanoff. 1972. *Pathogenesis of the Avian Embryo: an Analysis of Causes of Malformations and Prenatal Death*. John Wiley and Sons, New York.

Santolo, G. M., and J. T. Yamamoto. 1999. Selenium in blood of predatory birds from Kesterson Reservoir and other areas in California. *Journal of Wildlife Management* 63:1273–1281.

- Scheuhammer, A. M., A. H. K Wond, and D. Bond. 1998. Mercury and selenium accumulation in common loons (*Gavia immer*) and common mergansers (*Mergus merganser*) from eastern Canada. *Environmental Toxicology and Chemistry* 17:197–201.
- Schuler, C. A., R. G. Anthony, and H. M. Ohlendorf. 1990. Selenium in wetlands and waterfowl foods at Kesterson Reservoir, California, 1984. *Archives of Environmental Contamination and Toxicology* 19:845–853.
- Skorupa, J. P. 1998. Selenium poisoning of fish and wildlife in nature: lessons from twelve real-world examples. Pages 315–354 *in* W. T. Frankenberger, Jr., and R. A. Engberg, editors. *Environmental Chemistry of Selenium*. Marcel Dekker, New York, New York.
- Skorupa, J. P., and H. M. Ohlendorf. 1991. Contaminants in drainage water and avian risk thresholds. Pages 345-368 in A. Dinar, and D. Zilberman editors. *The Economics and Management of Water and Drainage in Agriculture*. Kluwer Academic, Boston, Massachusetts.
- Spalding M. G., P. C. Frederick, H. C. McGill, S. N. Bouton, and L. R. McDowell. 2000. Methylmercury accumulation in tissues and effects on growth and appetite in captive great egrets. *Journal of Wildlife Diseases* 36:411–422.
- U.S. Department of the Interior (USDI). 1998. Guidelines for interpretation of the biological effects of selected constituents in biota, water, and sediment. U.S. Department of the Interior, National Irrigation Water Quality Program Information Report No. 3, U. S. Bureau of Reclamation, Denver, Colorado.
- Wayland, M., A. J. Garcia-Fernandez, E. Neugebauer, and H. G. Gilchrist. 2001. Concentrations of cadmium, mercury and selenium in blood, liver, and kidney of common eider ducks from the Canadian Arctic. *Environmental Monitoring and Assessment* 71:255–267.
- Wiener, J. G., D. P. Krabbenhoft, G. H. Heinz, and A. M. Scheuhammer. 2003. Ecotoxicology of mercury. Pages 409–463 *in* D. J. Hoffman, B. A. Rattner, G. A. Burton, Jr., and J. Cairns, Jr. *Handbook of Ecotoxicology*. Lewis Publishers, Boca Raton, Florida.
- Yamamoto, J. T., G. M. Santolo, and B.W. Wilson. 1998. American kestrels (*Falco sparverius*) fed selenomethionine and naturally incorporated selenium. *Environmental Toxicology and Chemistry* 17:2494-2497.

Table 1. Selenium concentrations in  $\mu g/g$  dry weight (mean  $\pm$  standard error) in adult California gulls, their eggs, food, water, and sediment collected at Antelope Island, Hat Island, and Great Salt Lake Mineral (GSLM) colonies located on the Great Salt Lake, 2006.

	Antelope Island	Hat Island	GSLM colony	d. f.	F-value P	
Male mass $(n = 19)$	728 <u>+</u> 21 A	769 <u>+</u> 20 A	629 <u>+</u> 13 B	2, 16	12.24 0.0006	
Female mass $(n = 14)$	640 <u>+</u> 32	635 <u>+</u> 58	619 <u>+</u> 29	2,11	0.13 0.88	
Se in adult liver $(n = 35)$	7.3 <u>+</u> 0.7	7.8 <u>+</u> 0.6	9.2 <u>+</u> 0.9	2, 32	1.85 0.17	
Se in adult blood <sup>1</sup> ( $n = 35$ )	13.8 <u>+</u> 1.8 A	16.0 <u>+</u> 2.0 A	25.1 <u>+</u> 7.9 B	2, 32	6.27 0.005	
Se in eggs $(n = 35)$	2.8 <u>+</u> 0.2	3.1 <u>+</u> 0.3	3.4 <u>+</u> 0.1	2, 32	1.76 0.19	
Se in brine shrimp $(n = 15)$	3.4 <u>+</u> 0.1 A	5.5 <u>+</u> 0.1 B	4.6 <u>+</u> 0.1 C	2, 12	181.65 0.0001	
Se in water $(n = 3)$	0.5	0.6	0.3			
Se in sediment $(n = 3)$ 0.4		0.4 0.5				

<sup>&</sup>lt;sup>1</sup> Means in rows not sharing the same uppercase letter are significantly different (P ≤ 0.05). based on the Duncan's Multiple Range Test.

Table 2. Selenium concentrations in  $\mu g/g$  dry weight (mean  $\pm$  standard error) in adult California gulls, their eggs, food, water, and sediment collected at Neponsett Reservoir, Hat Island, and Great Salt Lake Mineral (GSLM) colonies located on the Great Salt Lake, 2007.

	Neponsett Reservoir	Hat Island	GSLM colony	d. f.	F-value	Р
Male mass $(n = 20)$	673 <u>+</u> 27	760 <u>+</u> 15	636 <u>+</u> 86	2, 17	1.15	0.34
Female mass $(n = 16)$	555 <u>+</u> 12	601 <u>+</u> 29	591 <u>+</u> 24	2, 13	1.49	0.26
Se in adult blood ( $n = 36$ )	15.5 <u>+</u> 2.3 AB <sup>1</sup>	10.7 <u>+</u> 1.4 A	20.9 <u>+</u> 3.4 B	2, 30	3.79	0.03
Se in adult liver $(n = 35)$	8.3 <u>+</u> 0.7	7.2 <u>+</u> 0.4	9.3 <u>+</u> 1.0	2, 30	2.20	0.13
Hg in adult blood ( $n = 36$ )	1.3 <u>+</u> 0.3 A	3.0 <u>+</u> 0.3 B	3.0 <u>+</u> 0.6 B	2, 30	7.38	0.003
Hg in adult liver $(n = 36)$	2.4 <u>+</u> 0.6 A	5.6 <u>+</u> 0.7 B	4.2 <u>+</u> 0.9 A	B 2, 30	5.52	0.01
Se in brine shrimp ( $n = 10$ )	No data	4.5 <u>+</u> 0.2 A	3.9 <u>+</u> 0.2 B	1,8	5.47	0.05
Hg in bring shrimp ( $n = 10$ )	No data	0.6 <u>+</u> 0.1	0.4 <u>+</u> 0.02	1,8	2.87	0.13

<sup>&</sup>lt;sup>1</sup> Means in rows not sharing the same uppercase letter are significantly different (P ≤ 0.05). based on the Duncan's Multiple Range Test.

Appendix 1. California gulls collected on 5/2/06 at the Great Salt Lake Mineral Colony (F= female, M = male, AL = active layer or female that has a large developing egg inside her, g = grams, mm = millimeters, L = length, H = height, and ww = wet weight,  $\mu g/g$  = micrograms of selenium per gram of tissue).

Sample	Sex	Mass (g)	Wing (mm)	Body (mm)	Head length	Bill L×H		Food in esophagus g (ww) Contents	Se (µg/ Blood	g) dry weight Liver
Cg-01	F-AL*	666	380	496	100	18×16	4.9	100% brine fly larvae	17	6.7
Cg-02	M	656	397	499	111	22×11	8.9	100% brine fly larvae	28	12
Cg-03	F	544	371	500	99	18×15	0.1	2 cori×ids	32	9.9
Cg-04	F-AL	697	384	490	98	18×15	9.1	100% cori×ids	37	13
Cg-05	M	633	370	450	99	18×14	0.0		13	6.1
Cg-06	M	635	395	527	109	21×17	0.0		18	7.5
Cg-07	M	644	379	495	111	23×16	5.7	100% brine shrimp	5	3.9
Cg-08	F-AL	579	399	495	99	18×14	5.9	100% brine shrimp	33	11
Cg-09	F-AL	542	385	475	98	17×15	1.0	100% brine shrimp	31	11
Cg-10	F-AL	687	375	495	101	17×16	6.4	(60% brine shrimp, 35% cori×ids, 5% midges)	25	8.6
CG-11	M	579	395	500	107	19×17	0.0		37	12

Appendix 1 (continued). California gulls collected on 5/4/06 at the Antelope Island colony (F= female, M = male, AL = active layer or female that has a large developing egg inside her, g = grams, mm = millimeters, L = length, H = height, W = width, ww = wet weight,  $\mu g/g$  = micrograms of selenium per gram of tissue).

Sample	Sex	Mass (g)	Ling (mm)	Body (mm)	Head length	Bill L×H	g (ww)	Food in esophagus  Contents	Se (µg/g Blood	g) dry weight Liver
A1	M	674	416	674	115	21×18	0.0		7.7	5.3
A2	M-subadult	787	380	510	108	20 ×14	0.0		20	6.9
A3	M	663	400	520	111	21×16	3.3	100% brine shrimp	19	9.5
A4	F-AL	665	385	490	100	19×14	13.9	100% brine shrimp	22	13
A5	M	731	404	500	107	23×16	157.0	100% carp carcass	14	6.1
A6	M	761	400	518	112	22×17	3.0	100% brine shrimp	25	9.9
A7	F	755	406	505	102	19×15	15.9	100% brine shrimp	13	6.0
A8	F	526	380	478	98	18×15	3.0	100% brine shrimp	13	6.7
A9	F	640	405	498	98	20×16	7.9	100% brine shrimp	7.7	4.0
A10	F-subadult	590	388	483	103	21×15	1.5	100% brine shrimp	8.8	6.5
A11	M	688	395	506	113	23×17	16.3	100% earthworms	10	6.9
A12	F-AL	669	386	490	96	18×10	1.2	100% brine shrimp	6.4	6.8

Appendix 1 (continued.). California gulls collected on 5/9/06 at the Hat Island Colony (F= female, M = male, g = grams, mm = millimeters, L = length, H = height, W = width,, ww = wet weight,  $\mu g/g$  = micrograms of selenium per gram of tissue).

Sample	Sex	Mass (g)	Wing (mm)	Body (mm)	Head length	Bill L×H	<u> </u>	Food in esophagus contents	<u>Se (μg/g</u> Blood	g) dry weight Liver
H1	M	806	395	478	112	22×18	16.1	100% brine shrimp	12	6.3
H2	?	767	395	496	109	22×17	27.1	100% brine shrimp	29	13
НЗ	F	693	382	480	105	20×15	6.4	100% brine shrimp	8.5	5.9
H4	M	767	400	520	109	20×16	33.7	100% brine shrimp	15	6.8
H5	M	854	394	520	107	21×17	24.3	100% brine shrimp	15	6.1
Н6	M	657	410	515	109	20×17	5.3	100% brine shrimp	17	8.4
H7	M	813	395	533	109	21×16	13.5	100% brine shrimp	16	9.3
Н8	F	578	360	505	99	18×15	0.3	100% brine fly larva	22	8.6
Н9	M	784	402	521	109	21×16	14.2	100% brine shrimp	18	8.6
H10	M	709	377	536	110	20×17	7.1	100% brine shrimp	25	9.3
H11	M	737	397	519	109	20×17	41.4	100% brine shrimp	8.1	5.7
H12	M	794	386	526	115	23×19	30.8	100% hot dogs	6.3	5.6

Appendix 2. California gulls collected on 5/7/07 at the Great Salt Lake Mineral Colony (F= female, M = male, g = grams, mm = millimeters, L = length, H = height, W = width,, ww = wet weight,  $\mu g/g$  = micrograms of selenium per gram of tissue).

Sample Sex	Body M g (ww)		Liver Mass g (ww)	Mass	Food in		ug/g) dry Blood	weight) Liver	Hg (u	g/g) dry v Blood		
GSLM -01	F-AL	662	26		9	100% brine shri		9.9	7.3		3.4	6.31
GSLM -02	F	562	14		4	100% brine shri	mp	13	6.8		6.02	9.94
GSLM -03	M	746	26		25	100% midge lar	va	28.3	14		2.3	3.1
GSLM -04	M	761	24		21	garbage (bread)		11	6.2		0.63	0.6
GSLM -05	M	741	20		46	garbage (bread)		21.8	7.4		1.0	1.0
GSLM -06	M	740	24		15	100% midge lar	va	28.9	9		3.2	5.03
GSLM -07	M	680	17		24	100% midge lar	va	13	7.2		0.72	0.92
GSLM -08	F	563	19		18	100% midge lar	va	17	11		1.1	1.3
GSLM -09	F	578	18		0.5	100% brine shri	mp	45.7	15		7.61	9.59
GSLM -10	M	736	23		15	100% brine shri	mp	38	14		3.7	6.3
GSLM -11	M	745	27		21	100% brine shri	mp	8.7	7		3.1	3.6
GSLM -12	M	645	26		14	100% brine shri	mp	16	6.5		3.5	3.3

Appendix 2 (continued). California gulls collected on 5/9/07 at the Hat Island Colony (F= female, M = male, g = grams, mm = millimeters, L = length, H = height, W = width,, ww = wet weight,  $\mu g/g$  = micrograms of selenium per gram of tissue).

Sample	Sex	Body mass g (ww)	Liver mass g (ww)	Mass	Food in crop Contents		g/g) dry weight) d Liver		g/g) dw) Liver
HAT -01	M	716	27	37	100% brine shrimp	23	9.7	3.4	6.57
HAT -02	M	722	20	28	100% brine shrimp	7.1	7.7	3.3	8.92
HAT -03	M	822	32	28	100% brine shrimp	13	8.8	3.4	4.6
HAT -04	M	789	17	85	garbage (bread)	13	5.9	4.3	5.27
HAT -05	F	635	16	45	100% brine shrimp	12	6.5	3.5	5.95
HAT -06	M	745	16	9	100% brine shrimp	4.8	4.7	0.56	0.77
HAT -07	F	612	20	27	90% garbage, 10% beetles	s 9	6.1	2.8	5.95
HAT -08	M	738	16	17	100% brine shrimp	7	6.7	2.6	3.8
HAT -09	F	673	16	16	100% brine shrimp	5.3	6.6	2.3	5.26
HAT -10	F	582	18	13	100% brine shrimp	12	8.3	3.5	6.3
HAT -11	M	788	21	57	100% brine shrimp	14	7.4	2.6	3.6
HAT -12	F	501	13	3	100% brine shrimp	8.7	8.3	3.3	9.8

Appendix 2 (continued). California gulls collected on 5/11//0 at the Neponset Reservoir Colony , Rich County, Utah Colony (F= female, M = male, g = grams, mm = millimeters, L = length, H = height, W = width,, ww = wet weight,  $\mu g/g$  = micrograms of selenium per gram of tissue).

Sample	Sex	Body mass g (ww)	Liver mass g (ww)	Mass	Food in crop Contents	Se (ug/ Blood	g) dry weight) Liver	Hg (ug Blood	
NET 01	E	556	16	2	1000/ James I flat lame	21	0.9	0.2	0.2
NET -01	F	556	16	2	100% damsel fly larva	21	9.8	0.2	0.3
NET -02	F	553	18	19	100% garbage (bread)	13	6.1	2.2	3.0
NET -03	F	550	18	27	100% garbage (bread)	10	8.1	0.2	0.36
NET -04	M	650	16	0	nothing	22.3	13	1.5	2.0
NET -05	F	560	16	43	90% caterpillars,10% bea	etles10	7.9	1.4	3.6
NET -06	M	612	16	0	nothing	5	5.6	0.75	0.93
NET -07	F	580	22	34	95% caterpillars, 5% bee	tles 24.8	8.2	1.4	5.93
NET -08	M	637	20	1	??	32.2	12	2.7	4.6
NET -09	F	492	17	0	nothing	8.2	6.8	0.21	0.32
NET -10	F	596	20	8	??	12	7.2	0.35	1.1
NET -11	M	764	17	7	2 bird leg bones	13	5.6	3.2	4.9
NET -12	M	700	17	0	nothing	15	8.7	1.0	2.3

Appendix 3. Selenium concentrations ( $\mu$ g/g dry weights), mass, size, developmental stage (H-H), presence of a viable embryo, and presence of an embryo with a visible defect. Eggs were collected at random from 3-egg clutches at the Great Salt Lake Mineral colony on May 15, 2006.

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Sample	e Se (με	g/g) <u>Mass</u>	(g)	Length	Width	Volum	е Н-Н	Viable	Defects?	
-		Whole egg	Without shell	(mm)	(mm)	(ml)		embryo?		
M-1	3.6	69.5	62.9	66.1	46.8	65.0	37	YES	NO	
M-2	3.0	70.1	63.9	69.7	45.0	63.9	37	YES	NO	
M-3	2.6	59.3	53.9	65.6	44.0	60.0	44	YES	NO	
M-4	3.2	70.3	62.8	68.1	48.1	76.5	44+	YES	NO	
M-5	4.1	64.0	58.2	64.5	46.4	-	45+	YES	NO	
M-6	3.7	69.7	62.9	68.9	43.6	70.8	44+	YES	NO	
M-7	2.7	63.6	52.1	67.9	46.0	63.7	42	YES	NO	
M-8	3.2	65.8	55.0	63.7	47.6	65.2	40	YES	NO	
M-9	-	67.2	61.0	67.2	47.0	68.6	39	YES	NO	
M-10	3.5	62.3	53.8	61.7	47.2	59.4	38	YES	NO	
M-11	4.3	70.6	64.1	63.5	47.4	66.6	44+	YES	NO	
M-12	3.3	62.6	54.3	65.2	47.0	-	45	YES	NO	
M-13		60.5	54.0	66.7	45.6	-	45	YES	NO	
M-14		65.8	56.5	64.0	46.8	61.4	38	YES	NO	
M-15		67.2	60.7	65.3	45.1	58.0	36+	YES	NO	
M-16		58.7	53.2	61.6	46.3	-	45	YES	NO	
M-17		65.9	57.5	68.7	45.9	56.8	-	NO	?	
M-18		66.4	59.8	64.4	46.1	54.8	39	YES	NO	
M-19		73.1	66.6	68.1	48.8	71.8	43	YES	NO	
M-20		66.3	59.7	64.1	46.9	-	44	YES	NO	
M-21		67.4	56.8	63.1	48.4	81.3	45	YES	NO	
M-22		68.0	59.9	66.0	46.1	78.0	37	YES	NO	
M-23		68.1	61.4	66.2	47.3	61.8	39	YES	NO	
M-24		54.8	47.0	58.3	47.7	-	44+	YES	NO	

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Appendix 3 (continued). Se concentrations ( $\mu g/g$ , dry weights), mass, size, developmental stage (H-H), presence of a viable embryo, and presence of an embryo with a visible defect. Eggs were collected at random from 3-egg clutches at the Hat colony on May 25, 2006.

Sample	Se (µg	g/g) <u>Mass</u>		Length	Width			Viable	Defects?
		Whole egg	Without (mm) shell	(mm)	(mL)		embryo	)?	
H-1	-	68.5	61.6	64.2	45.0	56.4	8	YES	NO
H-2	3.4	65.5	55.4	65.6	45.6	65.5	40	YES	NO
I-3	2.1	66.5	55.0	69.1	46.3	-	45+	YES	NO
H-4	3.4	66.0	57.8	63.5	46.1	64.4	41-42	YES	NO
H-5	3.3	66.5	59.2	67.0	44.7	62.1	37	YES	NO
H-6	2.8	63.6	56.6	63.6	46.2	62.2	43-44	YES	NO
H-7	2.3	63.1	53.2	62.5	46.3	64.0	43-44	YES	NO
H-8	-	75.3	64.3	64.7	48.2	74.5	42-43	YES	NO
H-9	3.1	72.1	65.1	65.3	47.3	71.3	36+	YES	NO
H-10	2.8	64.9	58.2	63.2	46.5	62.9	43-44	YES	NO
H-11	3.2	63.2	56.5	64.3	44.7	63.8	38+	YES	NO
H-12	2.5	57.7	52.3	62.7	44.7	60.2	45	YES	NO
H-13	2.0	67.4	58.1	67.0	45.5	66.0	42-43	YES	NO
H-14		69.5	63.0	67.3	47.1	72.1	44+	YES	NO
H-15		70.1	61.7	64.2	46.2	67.2	33	YES	NO
H-16		60.0	54.7	63.8	45.9	64.0	38+	YES	NO
H-17		67.7	61.0	68.0	44.9	68.6	41-42	YES	NO
H-18		58.4	51.6	64.3	43.4	57.5	43-44	YES	NO
H-19		72.4	63.8	67.8	46.0	71.7	37	YES	NO
H-20		63.9	55.5	66.9	44.5	64.5	42-43	YES	NO
H-21		63.2	53.3	64.4	44.7	62.9	44+	YES	NO
H-22		60.8	51.2	63.2	45.2	59.9	45	YES	NO
H-23		76.6	65.3	69.8	47.9	77.3	45	YES	NO
I-24		66.6	59.1	67.1	44.5	66.7	42-43	YES	NO

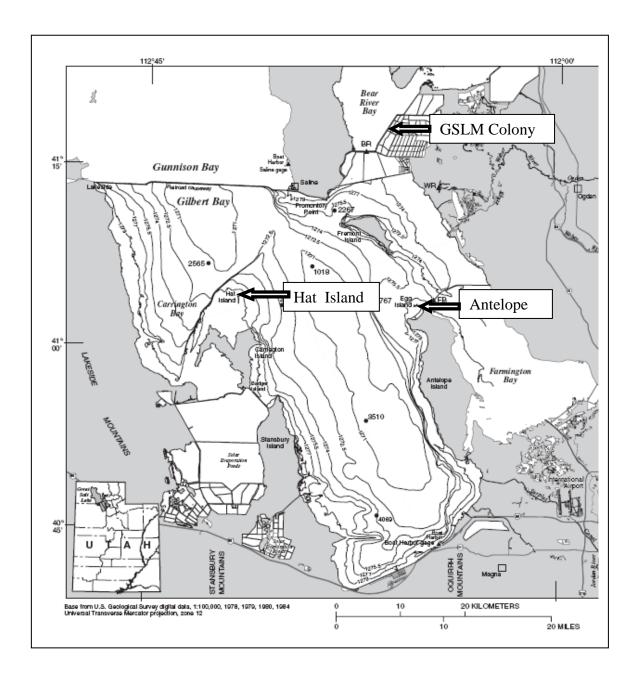
Appendix 3 (continued). Se concentrations ( $\mu g/g$ , dry weights), mass, size, developmental stage (H-H), presence of a viable embryo, and presence of an embryo with a visible defect. Eggs were collected at random from 3-egg clutches at the Antelope Island colony on May 23, 2006.

ample	e Se (µg	g/g) <u>Mass</u>	(g)	Length	Width	Volum	е Н-Н	Viable	Defects?
		Whole egg	Without (mm) shell	(mm)	(mL)		embryo	?	
<b>\</b> -1	3.2	62.6	53.4	64.0	45.4	62.9	42-43	YES	NO
<b>\</b> -2	3.0	63.4	55.0	62.2	46.9	66.0	44+	YES	NO
3	2.7	61.0	52.8	65.5	44.0	59.7	44+	YES	NO
4	4.1	57.8	52.6	62.4	44.3	-	45+	YES	NO
5	2.4	68.6	59.4	67.7	46.4	71.2	44+	YES	NO
<b>-</b> 6	-	61.5	54.2	60.6	47.2	-	45	YES	NO
<b>\</b> -7	2.1	58.1	51.4	62.1	45.4	62.5	38	YES	NO
<b>8</b> -2	2.6	61.3	53.5	66.5	44.6	61.5	41-42	YES	NO
<b>\-</b> 9	2.6	78.4	69.5	68.1	47.2	73.2	29	YES	NO
-10	2.4	82.0	71.8	67.1	48.3	-	23+	YES	NO
<b>-11</b>	2.4	68.9	62.6	69.2	45.3	65.7	41-42	YES	NO
-12	2.8	64.7	58.3	65.5	46.6	-	45+	YES	NO
-13		75.7	67.0	66.6	47.1	70.1	18	YES	NO
-14		72.1	62.6	64.2	47.2	68.8	39	YES	NO
<b>\</b> -15		63.7	55.2	62.0	45.8	62.6	42-43	YES	NO
-16		69.4	60.2	64.3	47.3	71.0	42-43	YES	NO
-17		63.9	54.5	65.6	45.5	-	45+	YES	NO
-18		69.2	62.6	65.7	46.4	66.9	37	YES	NO
-19		64.7	55.1	62.2	47.4	65.4	42-43	YES	NO
-20		67.0	55.4	63.0	47.2	66.8	45	YES	NO
-21		63.6	53.7	63.6	46.4	65.0	45	YES	NO
-22		65.8	59.6	66.4	46.5	69.8	44+	YES	NO
-23		71.6	64.5	66.8	45.6	68.6	35	YES	NO
-24		66.9	59.7	63.0	47.4	67.5	41-42	YES	NO

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Appendix 3 (continued). Se and Hg concentrations ( $\mu$ g/g, dry weights), mass, size, developmental stage (H-H), presence of a viable embryo, and presence of an embryo with a visible defect. Eggs were collected at random from 3-egg clutches at the Neponsett Reservoir colony on June 9, 2007.

r	Hg (µg/g)	Se (µg/g)	Mas Whole	SS (g) Length Without shell	Width (mm)	Volume (mm)	e H-H (ml)	Viable	Defects? embryo?	
-1	0.19	2.5	61.0	56.2	62.3	44.5	51	38	YES	NO
-2	0.16	2.5	68.5	63.2	64.3	46.1	61	33	YES	NO
-3	0.07	2.2	56.9	52.9	60.6	44.6	51	41-42	YES	NO
-4	0.37	2.7	68.4	63.0	62.6	47.1	62	33	YES	NO
-5	0.16	2.6	61.1	56.3	62.5	44.4	56	39	YES	NO
-6	0.24	2.4	58.5	54.2	66.0	43.1	53	34	YES	NO
-7	0.45	3.0	59.1	53.5	63.7	44.2	63		NO (rotten)	NO
-8	0.39	3.1	65.7	61.2	70.1	44.0	52	44+	YES	NO
-9	0.1	3.3	52.2	48.3	59.5	43.2	49	40	YES	NO
-10	0.7	3.0	66.5	60.7	65.9	45.2	56	34	YES	NO
-11	0.16	2.2	72.4	67.4	65.6	46.5	60		NO (infertile)	NO
-12	0.1	3.8	54.2	50.7	60.0	41.7			YES	NO



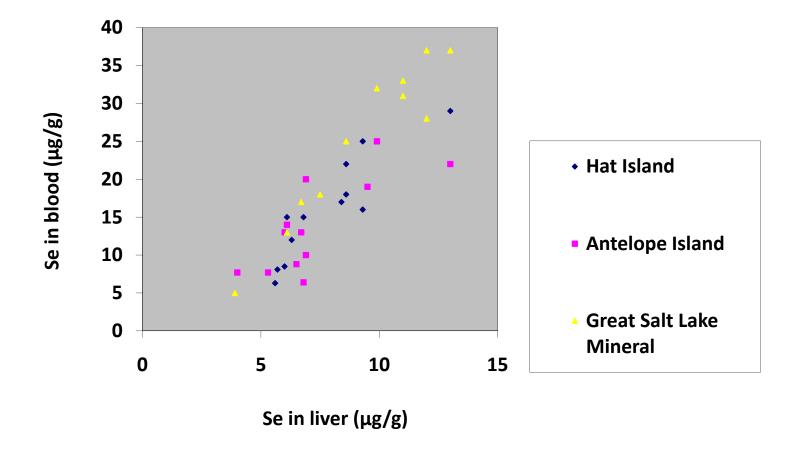


Figure 2. Relationship between a California gull's selenium concentration ( $\mu g/g$ , dry weight basis) in its blood and liver. Gulls collected from the three colonies are plotted separately.

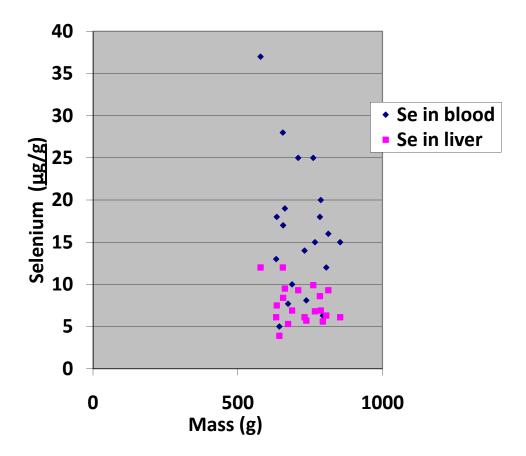


Figure 3. Relationship between the mass of male California gulls and the selenium concentration( $\mu g/g$ , dry weight basis) in their blood and liver.

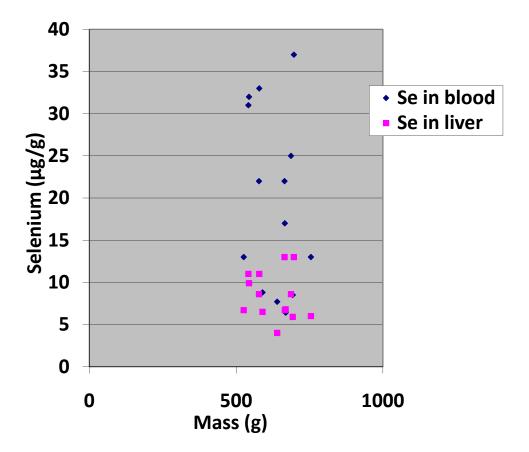


Figure 4. Relationship between mass of female California gulls and the selenium concentration ( $\mu g/g$ , dry weight basis) in their blood and liver.